

LEACHMAN RECYCLE AT SOLID WASTE LANDFILL
USING HORIZONTAL SECTION

By

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A DISSERTATION PRESENTED TO THE GRADUATE SCHOOL
OF THE UNIVERSITY OF FLORIDA IN PARTIAL FULFILLMENT
OF THE REQUIREMENTS FOR THE DEGREE OF
DOCTOR OF PHILOSOPHY

UNIVERSITY OF FLORIDA

1996

ACKNOWLEDGMENTS

The author wishes to thank his committee members for their assistance and review of this dissertation. A special thanks is offered to Dr. Lester Miller for his continued support and for providing the opportunity to work on a one-of-a-kind project. The author wishes to thank his fellow graduate students for their work and assistance during his studies.

Support for this project was provided by the Alaska County Department of Public Works. Thanks is offered to Ed Culpepper, John Carter, and the entire operating crew at ACHW. An extraordinary offer of appreciation is given to George Jennings for his work, much of which is described in this dissertation. This project was also funded by the U.S. Environmental Protection Agency and the Florida Center for Solid and Hazardous Waste Management.

The author is indebted to his parents for their continued support, emotionally and financially. Finally, the greatest thanks goes to Marlene, for her tremendous friendship, love, and support.

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Abstract of Dissertation Submitted to the Graduate School
of the University of Florida in Partial Fulfillment of the
Requirements for the Degree of Doctor of Philosophy

LEACHATE RECYCLE AT SOLID WASTE LANDFILLS
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May 1981

Chairman: W. Lamar Miller
Major Department: Environmental Engineering Institute

A method of leachate recycle known as horizontal injection was assessed. In this method, leachate is recirculated to a landfill through horizontal trenches or wells buried within the landfill. Potential advantages of a leachate recycle system of this nature include the ability to recycle leachate to the landfill without atmospheric exposure, the ability to control the point of leachate recycle and the resulting moisture distribution, and the ability to lessen impact of leachate recycle on daily landfill operations.

Analytical equations were developed to describe the expected shape of the saturated zone surrounding an injection line for both isotropic and anisotropic conditions. The equations were developed by treating an injection line as a

line source in a porous medium subject to a gravity drainage flow field. Capillary forces and dispersion were neglected. The equations allowed prediction of the horizontal and vertical spread of the saturated zone and the resulting injection pressure as a function of injection line length, flow, and waste hydraulic conductivity.

A horizontal injection leachate recycle system (HILRS) was constructed and operated at a Florida landfill. The system consisted of perforated pipes placed in shredded tire filled trenches and buried between lifts of compacted solid waste. The performance of 11 injection lines was evaluated. During a period of 19 months, 1.4 million gallons of leachate were recycled to the landfill through the injection system. Pressures and flows observed during the injection were recorded. Typical injection flow responses ranged from 0.008 to 0.027 gpm/ft per foot (water column) of applied pressure head in trenches with shredded tires as a drainage blanket. In trenches without tires, flow rates ranged from 0.004 to 0.008 gpm/ft per foot of applied pressure head. Injection at high pressures did result in some surface seeps from injection lines located near the surface of the landfill.

CHAPTER 1 INTRODUCTION

Leachate recycle, also referred to as leachate recirculation, is a method of managing leachate at solid waste landfills. This process involves the return of leachate intercepted by a landfill's liner and leachate collection system back to the landfilled solid waste. Benefits cited for leachate recycle include leachate volume management, leachate treatment, accelerated landfill stabilization, and enhanced gas production. Numerous pilot studies have been performed demonstrating these benefits (Folinski 1978, 1980; Jackis et al. 1979; Naisid et al. 1982).

Leachate recycle has been practiced at full-scale landfills in the United States and Europe (Barber and Davis 1984, Davis and Anderson 1988, Weiss 1987) but the results of these projects as regard to the performance and design of various methods of leachate recycle have been limited. The most recentSubtitle D regulations of the Resource Conservation and Recovery Act permit leachate recycle in properly designed landfills (Federal Register 1991), but the practice is discouraged by some state regulators as a result of potential threats from poorly operated systems. Further investigation into the mechanisms involved in various methods

of leachate recycle is necessary to allow safe and effective design of leachate recycle systems at landfills, and for the development of regulations and the transfer of this technology.

This dissertation discusses a leachate recycle process known as horizontal injection and reports its application at an operating landfill in Alachua County, Florida. In this process, leachate is pumped under pressure into horizontal injection lines buried at various depths within the landfill. This method of leachate recycle has a number of potential advantages over other methods including limitation of leachate exposure to the atmosphere, ability to control points of leachate application and leachate distribution, and minimal interference of leachate recycle operations on overall landfill operation.

In Chapter 2 of this dissertation, mathematical equations are developed to describe the shape of the saturated zone surrounding an injection line under steady injection conditions as a function of injected flow rate and the hydraulic conductivity of the medium. Isotropic and anisotropic conditions are considered. Chapter 3 presents information on the design, construction, and operation of the horizontal injection leachate recycle system (HILRS) at the Alachua County landfill, and discusses the results and performance of the system. Chapter 4 uses the information in Chapters 2 and 3 to outline construction and operation

guidelines, and a design procedure for use of horizontal injection as a method of inhibiting vapors at operating conditions. A summary of the discussion and conclusions reached from the research, as well as areas of additional needed research, are presented in Chapter 8.

CHAPTER 2
SATURATED FLOW FROM A HORIZONTAL INJECTION
LEACHATE RECYCLE SYSTEM

Introduction

Horizontal injection leachate recycle involves the recirculation of leachate to a landfill through the use of horizontal recharge wells. The wells are buried within the waste, and are usually constructed as the waste is deposited. Horizontal injection leachate recycle systems (HI-LRS) represent a promising technology for leachate and solid waste management at landfills, but such systems are relatively untested and little information is available for their design. This chapter outlines analytical equations which describe the characteristics of steady state saturated flow of water surrounding a horizontal well resulting from the pressure of the injected water and gravity. This chapter was developed in response to, but independent of, a HI-LRS under evaluation at a lined landfill in Florida (Chapter 3).

The method outlined here to model the saturated flow of water which surrounds a horizontal injection well is based on the assumption that the well can be treated as a horizontal line source. The line source is subject to the effect of gravity in the vertical downward direction. A further assumption is made that water movement is not impeded in the

downward direction. This case occurs in lined landfills where a highly conductive collection layer drains leachate from the liner system.

Simulation of water flow from line sources buried in soils has been performed in the soil science field in a number of previous investigations (Garda 1935, Philip 1971, Techuan and Thomas 1973, Thomas et al. 1984). The primary driving force for water movement in these irrigation studies has been the capillary action of the soil. These analyses have simulated the unsaturated moisture profile resulting from capillary forces, and describe conditions of relatively low flow. The resulting solutions are generally complex, analytical equations.

The process outlined here to describe the saturated zone surrounding a horizontal injection well involves the assumption that the primary driving forces for moisture transport are the pressure of the injected water and gravity. If capillary forces and the effects of dispersion are neglected, a single analytical solution describing the shape of the saturated zone may be determined. The resulting solution is similar to the solution of groundwater exchange through a vertical well in a confined aquifer subject to a uniform field of flow (Bear and Jacob 1954). The solution of the uniform flow field analysis has also been applied to the development of capture zone equations in the management of pollution plumes in groundwater (Cvetkovic and Chiu-Pu 1987).

The unconfined flow field in the vertical well scenario is analogous to the gravity drainage flow field in this analysis. An advantage to the treatment of the flow system in this manner is that a solution for the case of an nonwetting material is readily obtained.

Unsaturated flow and capillary forces certainly play a role in solute transport under typical landfill conditions. In the scenario outlined, a capillary fringe of water would develop surrounding the saturated zone. The equations utilized here are not intended to provide a method of conducting complex water balance and moisture routing simulations for landfills with LRS. Refined models which incorporate saturated and unsaturated flow and which account for the complexities of multiple well systems are necessary. It is proposed in this chapter, however, that for purposes of establishing design boundaries for efficient and safe RC-LRS operation at maximum flow conditions, saturated flow hydraulics can be used to estimate requirements for injection well spacing, depth, pressure requirements, and leachate collection system design.

Spreading Requirements

A schematic of the flow system evaluated is presented in Figure 2-1. The horizontal injection well is located with its center at the origin and is represented by a line source. The well consists of some finite radius (r_w). Water is injected into the well at a flow rate of q , which represents the linear

flow rate (flow per unit length of well), under steady conditions, a saturated zone will form around the injection well, reaching some maximum height above the well, γ_{sat} . At γ_{sat} , the potential of a fluid particle to move away from the well is equal to the potential created by gravity as move toward the well, and a stagnation point develops. It is assumed that flow may continue beneath the well without resistance. The width of the saturated front at an infinite downward distance is $2\gamma_{\text{sat}}$. The distance from the origin to the edge of the saturated front at the elevation of the well is defined as γ_{sat} .

Flow in Isotropic Medium

The system of saturated flow from a horizontal well in an isotropic medium is solved by the method of superposition in which the solutions of two simple flow fields are superposed. The first field consists of flow from a radial line source and the second field is gravity drainage. The potential function for a radial line source is

$$\phi = \frac{K}{2\pi h} \ln(r) \quad (2-1)$$

where ϕ is the potential, K is the hydraulic conductivity of the medium, and r is the radial distance from the line source.

The potential field created by gravity may be considered saturated drainage at a unit gradient. It is also assumed to represent the gravity field under conditions where saturated conditions do not occur, but where capillary forces and

displacements are neglected. Therefore, in the gravity drainage scenario, the potential gradient is 1, and

$$\frac{\partial \phi}{\partial y} = 1, \quad \frac{\partial \phi}{\partial x} = 0 \quad (2-2)$$

The solution for the potential function for the case of gravity drainage may therefore be written as

$$\phi = y \quad (2-3)$$

The two flow scenarios may be combined using the method of superposition. The variable x is converted to axisymmetric coordinates ($r^2 = x^2 + y^2$),

$$\phi = -\frac{Q}{16kH} \ln(x\sqrt{x^2 + y^2}) + y \quad (2-4)$$

A stagnation point occurs at $r_{\text{max}} = 0$. At this point, a singularity of zero flow develops where the potential gradient in the vertical direction is zero. The value of r_{max} may therefore be determined by solving Eqs. 2-4 when $x = 0$, and $d\phi/dy = 0$. This results in

$$r_{\text{max}} = -\frac{Q}{16kH} \quad (2-5)$$

The conjugate function of the potential function is the stream function, which describes the flow paths of fluid particles from the injection well. The stream function, may

be determined for a horizontal injection well with the relationship

$$\psi = -\frac{qB}{2\pi} \ln r \quad (2-6)$$

such that,

$$\psi = -\frac{qB}{2\pi h} \tan^{-1} \left(\frac{R}{r} \right) = \psi \quad (2-7)$$

The boundary for the saturated bulb occurs where $\psi = 0$, and this relationship may therefore be determined as

$$\frac{R}{r} = \tan \left(\frac{2\pi h \psi}{qB} \right) \quad (2-8)$$

The value of R_{sat} may be determined by evaluating Eqn. 2-8 for conditions when r approaches $\rightarrow 0$

$$R_{sat} = -\frac{qB}{4h} \quad (2-9)$$

From Eqn. 2-8, the value of R_{sat} may be determined at the boundary where $r = 0$.

$$R_{sat} = -\frac{qB}{4h} \quad (2-10)$$

In an isotropic medium, the stream lines will be perpendicular to the lines of equal potential. An example flow net for the saturated flow conditions surrounding the horizontal injection well is presented in Figure 2-3.

Flow in Anisotropic Medium

Often it is necessary to account for anisotropic conditions in a porous medium. This is especially true in

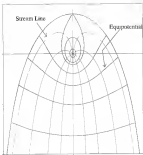


Figure 9-3. Equipotential and stream lines resulting from a horizontal injection well

situations which involve the flow of water is both the vertical and horizontal directions. Many times in natural soils, hydraulic conductivity in the horizontal direction will be greater than in the vertical direction a result of the manner in which sediments are deposited over time. This situation occurs in a landfill where waste is deposited and compacted in horizontal layers. Equations for the saturated zone surrounding a horizontal injection well in an anisotropic medium may be developed using the same method as the isotropic case. It is assumed the material is homogeneous and that axes of anisotropy correspond to the principal axes of the system. The governing equation for the case of steady plane flow is

$$K_x \frac{\partial^2 h}{\partial x^2} + K_y \frac{\partial^2 h}{\partial y^2} = 0 \quad 2-41$$

where K_x and K_y are the hydraulic conductivities in the x and y direction, respectively. The anisotropic problem may be solved by means of an equivalent isotropic flow region by an appropriate scale change. The method has been previously outlined by Bear and Dagan (1968). The solution of the potential function for a line source in an anisotropic medium using the method of an equivalent isotropic flow region was presented by Currier and Jaeger (1969). When combined with Eqs. 2-3 by the method of superposition, the resulting potential function for an injection well in an anisotropic medium is

$$\theta = \frac{Q}{4\pi k_y H \Delta T_e} \ln \left(\frac{R^2}{R_w^2} + \frac{R_w^2}{R_y^2} \right) + C \quad 2-12$$

Using the same techniques applied in the isotropic case, the equation of the saturated zone may be determined as

$$R = \frac{R_w}{18k_y} \cosh^{-1} \left(\frac{R}{R_w} \sqrt{\frac{k_y}{k_x}} \right) \quad 2-13$$

The parameters R_{max} , R_{min} , and R_{max} may be determined as

$$R_{\text{max}} = \frac{R_w}{18k_y \sqrt{k_x/k_y}} \quad 2-14$$

$$R_{\text{min}} = \frac{R_w}{18k_y} \quad 2-15$$

$$R_{\text{max}} = \frac{R_w}{18k_y} \quad 2-16$$

The distances R_{min} and R_{max} are a function of the hydraulic conductivity in the vertical direction only. The location of the stagnation point above the well, R_{max} , is function of hydraulic conductivity in the vertical and horizontal direction, and thus varies for different degrees of anisotropy. The effect of the degree of anisotropy on the shape of the saturated bulb is illustrated in Figure 2-8.

Injection Flow and Pressure Relationship

An additional parameter of interest is a horizontal injection well system is the injection pressure required for a given flow rate in a given set of well conditions. The

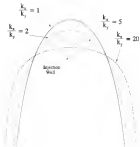


Figure 2-3. Effect of Anisotropy on Submerged Boundary

pressure head required, P_0 , may be defined for anisotropic conditions as

$$P_0 = \frac{Q}{2\pi K} \ln \frac{R}{r_w} \quad (2-17)$$

and for anisotropic conditions as

$$P_0 = \frac{Q}{2\pi K_0 K_{xy}} \ln \frac{R}{r_w} \quad (2-18)$$

where r_w is the effective well radius and R is the radius of influence of the line source. In the mathematical relationship for a line source, the radius of influence continues outward towards infinity. It is therefore necessary to assume some radius at which the effect of the line source is negligible. In the case of a horizontal injection well, this may be viewed as the distance beneath the well at which the effect of the well is no longer observed. The radius of influence may be selected in the same manner as typical well hydraulics using empirical or semi-empirical estimates (Bear 1979). These estimates often vary widely and P_0 should be evaluated for a range of R values.

Leachate Flow to the LCH

The flow of leachate into the landfill bottom drainage layer is a critical part of the design of the leachate collection system (LCH). If the LCH was located at an infinite distance beneath an injection line, then the flow into the LCH would be equal to the injected flow rate.

However, the closer the LCR is located to the injection line, the greater the flow rate per unit area. The flow rate for a given area of the flow net may be determined from the difference between the stream function lines, which may be estimated using Eqs. 3-7.

Discussion

The steady state equations presented here allow an estimation of the shape of the saturated flow zone that results from a horizontal injection well in a landfill which drained by a LCR. Graphs presenting the change in V_{sat} , V_{unsat} , and S_u as a function of the linear flow rate and the hydraulic conductivity are presented in Figure 3-4, 3-5, and 3-6 respectively.

The greatest limitation to the application of this analysis is the neglect of unsaturated flow conditions. Unsaturated flow does occur in a landfill, and is the predominant means of moisture transport for low moisture flow conditions. In the case of a horizontal injection well under pressure at steady state, the forces of pressure and gravity will be the dominant transport mechanisms in the area surrounding the wells. A zone of moisture transport will occur in the area surrounding the saturated zone as a result of capillary forces, and may affect the shape of the zone. At larger pressures, the potential created by hydrostatic pressure should be much greater than from capillary potential.

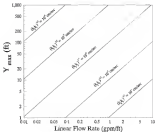


Figure 3-4: Y_{max} as a Function of Linear Flow Rate

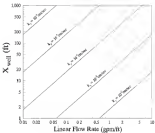


Figure 2-5. X_{well} as a Function of Linear Flow Rate

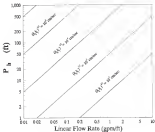


Figure 3-6. Pressure Versus Linear Flow Rate ($k_0/r_p = 1.0$)

The above equations were developed to allow design under maximum flow conditions which might be encountered in RI-LRS operation. For example, seepage into the LRS will be at a maximum when a saturated zone develops surrounding an injection well and reaches into the LRS. Although planned LRS operation may not reach steady state, the maximum possible conditions should be designed for to accommodate future operating conditions.

Conclusion

Equations were developed to describe the shape of the saturated zone surrounding a horizontal injection well under steady state conditions. The injection well was treated as a line source subjected to a gravity drainage flow field, and capillary forces and dispersion were neglected. Solutions were presented for isotropic and anisotropic cases. The predominant driving forces of water movement were assumed to be the pressure of the injected water and gravity. The method of images outlined here is proposed as a method to establish design criteria for the application of a RI-LRS at full-scale landfills. The equations allow a means to design injection well spacing, and the depth from the surface at a given flow rate. An estimate of the pressure head, which is necessary to properly size the pump and piping system, can be obtained as well.

CHAPTER 5
LEACHATE RECYCLE USING GEOTECHNICAL ENGINEERING
AT A FLORIDA LANDFILL

Introduction

Leachate recycle involves the recirculation of leachate collected from a lined landfill back through the landfilled waste. This practice provides leachate volume management and offers the potential to accelerate the decomposition of biodegradable waste in a landfill. Leachate recycle has been successfully demonstrated in pilot-scale studies to achieve waste stabilization (Lockie et al. 1979, Fokland 1980), and has been practiced at a number of full-scale landfills with some positive results (Barber and Meria 1984).

The recycle of leachate in a landfill may be performed by a number of methods including spray irrigation, surface application, and direct recharge to the interior of the landfill. Vertical recharge wells have been utilized to recycle leachate, but the resulting distribution of moisture through the waste is questionable. Horizontal injection wells offer the potential to distribute moisture over large areas of a landfill, while avoiding the problems associated with leachate exposure to the environment such as migration of odor and aerosols. Horizontal injection wells minimize

interference with normal landfill operations and vehicle and equipment traffic. The concept of leachate through horizontal distribution systems has been briefly mentioned in other studies (Smith and Anderson 1983), but an detailed study of their construction and performance has been reported.

This paper presents the results of a horizontal injection leachate recycle system (HILRS) at a lined landfill in Florida. The construction, operation, and performance of eleven injection lines, are reported for a period from February, 1983 through August, 1984. The information is presented to facilitate the use of this technology at other landfill sites, and provides information relevant to the construction and operation of HILRS and the design of such systems. The effect of the HILRS on a number of additional landfill parameters, including leachate generation and composition, waste decomposition, landfill temperature, and gas production, were monitored at this research site, and will be reported elsewhere.

Methods

Site Description

The horizontal injection leachate recycle study was performed at the Alachua County Southwest Landfill (ACSWL) in North Central Florida (Figure 3-1). The 37-acre (11-hectare) landfill is lined with a composite liner and is equipped with a leachate collection and removal system. Leachate drains

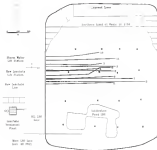


Figure 1-1. Plan View of ACNSL with Injection Lines

from the landfill to a lift-station and be pumped to a leachate equalization-treatment plant. The landfill includes an overline and line-precipitation leachate treatment system. Treated leachate is hauled off-site for disposal at a waste water treatment plant. An infiltration pond leachate recycle system (LRS) was operated beginning in 1978 (Tommaso et al. in press a) as part of a project involving the accelerated stabilization of the landfilled waste (Tommaso et al. in press b).

The site receives an average of 40 inches of rain annually. Approximately 500 tons per day of waste were deposited in the lined landfill unit at the time of this research. Apparent waste compaction is approximately 18% to 19% lb/yd^3 . Cover soil consists of sand mixed on site. Waste is deposited in a set of four lifts which are placed in an east-west fashion and which slope upward to the north against the previous set of lifts.

LRS Construction

The construction of horizontal injection lines (HIL) began in December 1981 on a series of four lifts constructed north of the infiltration pond LRS. These injection lines were placed in this series of lifts, the first on top of lift 1, the second on top of lift 2, and the third on top of lift 3. The result was a stepped positioning in which the first injection line was the northernmost and positioned at the lowest elevation, and the third injection line was the

southernmost and at the highest elevation. Upon completion of one set of injection lines, another set of three lines was placed in the next series of lifts in a similar fashion. The resulting cross section is shown in Figure 3-3, displaying the position of the injection lines in place at the end of August 1984, eleven of which were used and discussed here.

The characteristics of each HIL are listed in Table 3-1. The length and location of each injection line was dependent on the available area of lift. The design pipe diameter for the leachate recycle piping system was 3-inches, and this diameter was used for the injection lines as well. Holes were drilled in the injection pipe before installation. The hole size and spacing in the original injection line were designed to distribute flow evenly throughout the trench. It was felt that resistance to injected leachate would most likely result from the solid waste surrounding the trench, and the trench would fill with leachate under pressure, obviating the need for distributed flow from the pipe. The later injection lines (HILs 3-11) were therefore more designed (with additional holes) to accommodate for potential sealing and clogging.

PVC pipe was selected for the majority of the injection lines because of ease and quickness of installation. The PVC pipe was affixed with cement during installation. A high density polyethylene pipe was installed for HIL 9 and was disconnection prior to installation. The length of each injection line placed was dictated by the length of individual

Table 1-1. 34 Section Values Characterization

34S. Section	Date at 34S. Section	Pipe Material	Pipe D.D. (in.)	Length of Section (ft.)	Bole Size (in.)	Bole Spacing	Distance Between Boles	Average Disturbance (CL, 100%)
1	Dec 82	PVC	8.25	450	6.25	6	Tube Borehole	87
2	Dec 82	PVC	8.25	750	6.25	6	Tube Borehole	104
3	May 83	PVC	8.25	1400	6.25	6	Tube Borehole	113
4	Jul 92	PVC	8.25	300	6.25	6	Tube Borehole	87
5	Jul 92	PVC	8.25	120	6.25	6	Tube Borehole	104
6	Aug 92	PVC	8.25	140	6.25	6	Tube Borehole	113
7	Dec 91	PVC	8.25	160	6.25	6	Tube Borehole	87
8	Jan 93	PVC	8.25	120	6.25	6	Tube Borehole	104
9	Feb 93	PVC	8.25	140	6.25	6	Tube Borehole	113
10	Apr 93	PVC	8.25	1400	6.25	6	Bole	87
11	Apr 93	PVC	8.25	120	6.25	6	Bole	104

* The last spacing is hole per a 2 ft average.

* One hole every two feet.

* The holes every two feet.

lifts at the time of installation. Injection lines 1 and 2 were constructed to cross the entire bottom length of the landfill (450 ft and 180 ft, respectively). This necessitated building the lines in phases because the entire lift lengths were not constructed during one time period. The practice of phased construction required precise coordination of waste placement and reconstruction of part of the trench. To minimize these disturbances, the remaining injection lines were constructed in one phase, utilizing the length of lift available.

Construction of the injection trenches for lines 3 through 11 were conducted with a backhoe, which was used to excavate a trench down the center of the lift from west to east. The trench for MTL 1 was constructed using a bulldozer, which resulted in a much larger trench volume. The backhoe trenches were approximately three feet square, and the excavated waste and cover material were placed to the side of the trench. Shredded tires were used as drainage material in injection lines 1 through 8. The tires were placed at the bottom of the trench to a depth of approximately 1 foot. Tires were not placed in lines 10 and 11 for comparison of injection performance.

The pipe was placed in the trench and grout. An additional one to two feet of shredded tires (if used) were placed on top of the pipe and spread evenly. The terminal end of the pipe was capped. Non-perforated pipe was installed in

the first 20 ft of trench and as tires were placed in this area of the trench to prevent channeling along the pipe. This trench area was covered only with waste. The injection lines were connected to non-perforated riser pipes up the side slope of the landfill. The injection line area of the trench was covered with excavated waste and compacted using a progression of equipment, from light to heavy. An effort was made to keep the trenches covered on all sides with waste, and not cover sand, which would lead to short circuiting of injected leachate.

Leachate injection lines were connected to a main LBL manifold line which ran along the western outer landfill berm, to the DFF and two submersible oblong steel well pumps (1 hp and 3 hp) in the raw leachate equalization basins. Each EIL was valved separately to allow injection into any desired lateral. The pumps were valved to allow independent and parallel operation.

System Operation and Monitoring

Two parameters were assessed to assess the hydraulic performance of the system, leachate flow and pressure. Flow was measured using a paddle-wheel flow meter transducer (analog output) installed in the main manifold line on the berm. Pressure was measured with an analog output pressure transducer located in the manifold line on the landfill berm 50 ft before the first EIL. Analog signals were recorded by

data loggers in the field which were periodically downloaded to a portable computer.

Injection Schedule

The application of leachate to various injection lines, and the frequency and rate of application, were dictated by a number of factors. These factors included research goals of monitoring the effect of the LSI on the landfill system, the volume of water available for injection, and the schedules of the landfill operating crew. The injection of leachate was suspended at times to allow the landfill system to be monitored and for the repair and maintenance of the SI-LSI. All of the injection data reported here involved leachate ¹ recycled into one injection line at a time.

Initial injection tests were conducted during February through April 1993, using injection lines 1 and 2. During the initial test period, leachate was injected for prolonged periods to observe the physical operating constraints of the system. The next round of injection tests began in September 1993, and from that time leachate was injected on a daily basis, at a maximum of 40,000 gallons per day. This limit was an arbitrary level established by the landfill's permitting authority. The schedule of injection during the period from September 1993 to January 1994 consisted of rotating injection into lines 1 through 4 at approximately 40,000 gallons per day for two or three days into each line. In the spring and summer of 1994, leachate injection began into lines 1, 2, 3,

18, and 19, with continued limited injection into lines 1 through 4.

Excavation and Backfill

WIL Construction

Construction of the WIL was performed by landfill operators using typical landfill equipment. The most noticeable difficulties encountered in WIL construction were timing of WIL construction and odors resulting from uncovered waste. Because the injection lines were constructed between lifts of waste, the available construction time was often only a matter of days. This was especially true in the cases where the end phase of the lower lift, the beginning phase of the upper lift, and the inlet section of an WIL coincided. In these cases, WIL were often constructed in the presence of other landfill vehicles. The location of the trench on the surface of the lift was determined by considering the location of the linerrock roads, the proximity to large cover soil layers, the proximity to vertical gas wells, and the slope available for the backhoe to excavate the waste. Although a rubber-tired backhoe was used at MSWLF, a back-hoe on track would have allowed more flexibility for WIL placement in areas of steeper slopes.

The trenches were excavated in a manner designed to avoid direct contact with thick lenses of cover material which would result in short-circuiting. Complete avoidance of these lenses was impossible. Covering the trench with solid waste

and compensating the area above the trench required constant attention of the landfill operator to ensure sufficient waste was on top of the line before heavy equipment was used. The shredded tires did act as adequate bedding and cover for the pipe in this process.

Water presented a problem when large areas of trench were left open for periods of time greater than 12 hours. This was controlled by installing the drainage material and pipe during or immediately after trench excavation in such a manner as to ensure that no trench area was left open more than 12 hours.

The injection lines were checked for integrity prior to use. Of the first 11 injection lines, two were found to be obstructed. These RIL were used in the upper level and in the non-perforated pipe section. This area of the injection system was most vulnerable to damage from vehicles and heavy equipment because of the proximity of the pipe to the landfill surface and access roads. Injection lines on the lower levels were generally covered with large thicknesses of waste soon after completion, and traffic was not a problem for these. The proximity to the surface, did however, permit observation and repair of damaged pipe sections. The obstructions were the result of landfill equipment or trenches crushing pipe and both occurred near the major access road for solid waste delivery vehicles. After the initial spring 1992 test period, a blockage was discovered in RIL 1 at 480 ft during the installation of a temperature probe.

While a detailed analysis of the materials used in the HIL construction was not performed in this study, it should be considered in the design of such a system. The presence of landfill equipment and the occurrence of differential settlement create conditions that can cause pipe breaks. Differential settlement presents the greatest risk to pipe breakage in lines constructed over large thicknesses of waste. The greatest concern at MSWLF was a break in the pipe in the non-perforated riser section, because such a break would result in loss of use of that line. A pipe break in the perforated section of the pipe presented less of a concern in the HIL which used shredded-tire drainage material. In the event of a break, the tires would still allow some hydraulic transmission.

Landfill settlement did result in stress at the piping connections of the HIL and the HIL manifold. This resulted in some leaks in valves and connection. Future construction should include flexible hose connections to replace any rigid connections between pipes inside the landfill and pipes outside the landfill.

Landfill Leachate Volume

A total of 3,800,000 gallons of leachate were recycled using the HIL-180 during the period from February 1993 to August 1994. The total volume of leachate injected into each

Table 3.3. Injection Spacing Summary

Well	Start Injection	Total Hours of Injection	Total Volume Injected (gallons)
1	Feb 4, 83	313.3	2,344,380
2	Mar 27, 83	388.3	2,827,180
3	Apr 18, 83	84.4	278,800
4	Oct 12, 83	178.6	848,800
5	Oct 1, 83	187.8	888,800
6	Apr 27, 84	81.3	383,800
7	Mar 3, 84	303.3	1,898,480
8	May 6, 84	113.8	438,800
9	Jan 8, 84	71.3	138,800
10	Feb 6, 84	87.3	348,800
11	Feb 18, 84	83.8	378,800
Total	--	1718.4	7,814,180

line during this period is presented in Table 3-3. The greatest volume of leachate was injected into HIL 1, primarily as a result of the initial injection period in the spring of 1966. As a result of the length of this line, and the large branch volume, the initial conditions offered little resistance to flow. Consequently, greater volumes of leachate were accepted to HIL constructed on the lower and middle lifts relative to HIL on the top lifts. The operation of the injection system in these upper lifts resulted in occasional surface springs as the flow rates and pressures created by the leachate recycle pumps. The seeps occurred at points closest to the venturis and (later) of the perforated pipe section, and resulted from short circuiting through more conductive soil material or channels created by previous seep borings.

Pressure-Flow Relationship

The leachate flow rates and pressures in the main venturis were recorded for each injection run. A typical flow and pressure response occurring during an injection period is presented in Figures 3-4. The flow rate started at the highest value, and decreased towards a steady state. The pressure started at a low value and tended towards a higher steady value. The slight decrease in pressure at steady state conditions was the result of dropping leachate levels in the leachate storage tanks. The degree of this change and the time to reach a steady value were a function of the volume of leachate previously injected, the degree of saturation,

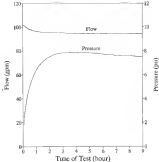


Figure 3-2. Typical Injection Flow and Pressure Response

resulting from any nearby injection lines, and the total storage volume in the trench.

The leachate flow and pressure responses were recorded for each EIL as a function of calendar time and cumulative pumping time, and these results are presented in Appendix A. To factor indifference to EIL length and to estimate the actual applied pressure at the inlet of the perforated section of an EIL, flow and pressure at the metering point were converted to linear flow rate (gpm per foot of injection line) and applied pressure head (feet of water column). The applied pressure was estimated from the pressure recorded in the line at the pressure gage, the average elevation of the line, and expected head loss from the pressure gage to the EIL inlet. Figure 3-4 presents the linear flow rate and applied pressure head for EIL 3, a line which received a total of 1.3 million gallons of recycled leachate. The data are presented as an average at 0.1 day time increments. The pattern of decreasing flow and increasing pressure at later time periods was observed in the long term trend results as well. The results are presented in Appendix A for each EIL.

The flow rate which occurs at a given applied pressure provides information regarding the performance of the system, its change over time, and is data needed for the design of similar systems. A parameter α is defined to represent the ratio of linear flow rate (gpm/ft) to the applied pressure head at the inlet of a EIL (ft w.c.).

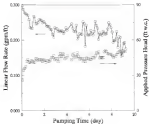


Figure 3-4. Linear Flow Rate and Applied Pressure Head versus Time² SIL 7

Table 3.3. Average rejection flow rates and applied pressure

Rejection rate (%)	Flow Rate (gpm)	Average Linear Flow rate (gpm/ft)	Average Applied Pressure Head (ft. water)	Average Rejection rate (gpm/ft/ft)
1	78.8	0.128	42.2	0.00343
2	71.8	0.092	38.4	0.00343
3	64.8	0.127	38.7	0.01003
4	77.4	0.215	41.8	0.00814
5	88.3	0.148	33.7	0.00814
6	78.4	0.147	33.8	0.01338
7	78.8	0.221	43.4	0.00509
8	77.8	0.148	35.9	0.00514
9	72.4	0.134	35.8	0.00897
10	72.7	0.202	43.8	0.00411
11	68.8	0.138	34.2	0.00448

The average results of flow rate, linear flow rate, applied pressure head, and α , are presented in table 3-3 for each RIL. The results represent averages for the entire period of injection, with exception of RIL 1 and 3, which do not include the data from the spring 1973 test. While a large volume of leachate was recycled during the spring 1973 test, a number of the data files were lost, and only one pump was installed during this time. Some pump clogging was observed which proved to provide erratic results in regard to the pressure and flow data (see Appendix A). The value of α as a function of cumulative pumping time is presented in Figure 3-2 for RIL 4 and 5, Figure 3-4 for RIL 3 and 6, and Figure 3-5 for RIL 8, 9, 10, and 11.

A number of observations were made regarding the pattern of α as a function of time. The RILs which resulted in the greatest measured flow to pressure responses were those nearest the landfill surface. These were the RILs in which the likelihood of surface springs was the greatest. The factors likely played a role in this phenomenon. First, it was realistic that the hydraulic conductivity was greater in this area because of smaller effective stress resulting from smaller depths of waste above these injection lines. In addition, the large pressures which occurred in the injection system at ACMT created upward waste uplift conditions so that the force of the injected leachate would create preferential channels. During injection into a surface RIL,

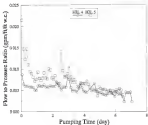


Figure 3-5. α versus Time; MIL 4 and 5

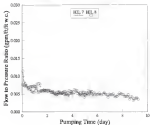


Figure 3-6. q versus Time: Well 7 and 8

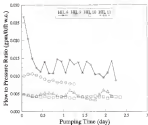


Figure 3-7. α versus Time²⁰ WEL 4, 5, 10, 11

leachate injection was observed to visibly uplift a limestone rock with the formation of a surface spring. The results in the upper hills were the most variable because of this phenomenon and because these lines were operated on a very intermittent basis.

As was previously observed with the flow and pressure relationships over time, the values of α decreased with time as appeared trend for all the hills. A number of potential reasons may be cited to account for this. The initial reduction in flow response was a result of the saturation of the surrounding waste and trench, and the filling of the available storage volume. A potential reason for long term reduction is the occurrence of clogging or fouling of the perforations in the pipe, and the trench itself. Another possibility which was visibly observed at the landfill was the back pressures of landfill gas in the injection lines. At times after an injection run, gas pressures as high as 100 inches of water column were recorded. The higher gas pressures, which would result from increased gas production and reduced void space in the near-saturated waste, would result in reduced relative permeability for water flow through the solid waste. To accommodate long term flow reduction, the ability to operate at higher pressures, and the ability to extract gas should be considered as part of system design.

The most representative flow responses for the injection lines were observed in Hills 4, 5, 7, 8, 10, and 11. These

lines were buried deep enough in the landfill to avoid surface seepage problems. The average α values in lines 4, 5, 7 and 8 ranged from 4.0881 to 5.0245 gpc/ft/ft w.c.. The average α values in lines 10 and 11 (as tire drainage material) were 0.0881 and 0.0894 gpc/ft/ft w.c.. The lojection system without wires still functioned, although at slightly larger pressures. Construction without tires made the installation process much easier and more rapid for the operator, but a pipe break in one of these branches would greatly reduce injection performance.

The range of values presented above represents the most typical conditions at ACNE as measured during the lojection experiments and the ones which should provide the best reference for the design of similar systems. It should be noted that the α values measured were for a specific flow range and extrapolation beyond this flow range may provide incorrect results. Future tests at the ACNE EI-LRI should include the performance of injection tests at lower flow rates for prolonged periods of time to measure the steady state response of the system.

Horizontal Conductivity Estimation

In a previous chapter (Chapter II), an equation was presented which related the injection pressure to flow rate for a horizontal well based on line source hydraulics, such that:

$$P_2 = \frac{q}{2\pi k_x k_y} \ln \left(\frac{R}{r_w} \right) \quad 3-1$$

where q is the linear flow rate, k_x and k_y are the hydraulic conductivities in the x and y directions, R is the radius of influence for the line source solution, and r_w is the effective radius of the well.

Although the lejection tests reported in this paper were not designed to determine the validity of this relationship, the data collected can be used to estimate the magnitude of the hydraulic properties of the waste. An additional fact that makes this exercise useful is that previous tests were conducted in the same landfill using an infiltration pond test.

The vertical hydraulic conductivity, k_y , was estimated from those experiments at 4×10^{-2} to 4×10^{-3} cm/sec. The lejection test data can therefore provide some estimate of the degree of anisotropy. The infiltration pond tests were performed on an area south of the RI-100. Exploratory borings in the two areas suggests that the waste composition was greater in the pond area than in the lejection area.

One difficulty encountered in the use of a line source solution is the necessity to assume a value for the radius of influence. Table 3-4 presents hydraulic conductivity estimates for the most representative RILs (4,5,7,8) for values of R/r_w of 100 and 1000. Hydraulic conductivity estimates are made for an isotropic assumption, and for anisotropic assumptions with k_y values of 4×10^{-3} cm/sec and

Table 3.4. Hydraulic Conductivity Estimates

RT ₀	$\frac{R}{\rho g \mu \eta (1/\text{Pa} \cdot \text{s})}$	R/r _c	$\frac{K}{\text{cm}^2/\text{sec}}$	$\frac{K}{\text{cm}^2/\text{day}}$
6	0.0001	100	2.88×10^{-4}	2.46×10^{-5}
		100	1.25×10^{-3}	8.46×10^{-5}
		100	4.00×10^{-3}	1.04×10^{-2}
		100	3.47×10^{-4}	3.07×10^{-5}
		100	1.40×10^{-3}	9.43×10^{-5}
		100	4.50×10^{-3}	3.36×10^{-2}
6	0.00035	100	3.45×10^{-4}	3.00×10^{-5}
		100	1.40×10^{-3}	1.19×10^{-4}
		100	4.40×10^{-3}	3.94×10^{-2}
		100	4.14×10^{-4}	4.14×10^{-5}
		100	1.40×10^{-3}	1.71×10^{-4}
		100	4.40×10^{-3}	4.26×10^{-2}
7	0.00035	100	3.33×10^{-4}	2.91×10^{-5}
		100	1.25×10^{-3}	8.41×10^{-5}
		100	4.00×10^{-3}	1.04×10^{-2}
		100	3.44×10^{-4}	3.04×10^{-5}
		100	1.40×10^{-3}	9.71×10^{-5}
		100	4.40×10^{-3}	3.31×10^{-2}
8	0.00035	100	3.34×10^{-4}	2.90×10^{-5}
		100	1.40×10^{-3}	9.71×10^{-5}
		100	4.40×10^{-3}	3.18×10^{-2}
		100	3.44×10^{-4}	3.04×10^{-5}
		100	1.40×10^{-3}	1.26×10^{-4}
		100	4.40×10^{-3}	3.14×10^{-2}

1000^{-1} cm/sec. The estimated hydraulic conductivities ($(k_{xy})^{(1)}$) were found to be more than one order of magnitude greater than the previous ACFM k_y estimates, indicating a definite anisotropic nature of the landfill material. The anisotropy at ACFM would result from the composition of the waste, the manner in which the waste was deposited and compacted, and the presence of sandy cover soil layers.

Summary and Conclusions

A horizontal injection leachate recycle system was successfully operated at a lined landfill in Florida. The data reported here included the construction and operation experience, and the hydraulic performance of the RI-LRS. Over a 18-month period, 7,900,000 gallons of leachate were recycled into 11 injection lines. Construction of the lines was performed by landfill operators with typical landfill equipment.

Applied injection pressure and leachate flow rates were monitored throughout all of the injection tests. The flow rate to applied pressure ratio was monitored, and decreased with time. Substantial reasons for this observation were clogging of injection lines and branches, and the large gas head-pressures observed in the lines. Substantial volumes of leachate would still be recycled to the landfill at these later times. Long term monitoring is needed to observe effects on flow rates after extended operation times. Typical values of flow response ranged from 0.005 to 0.007 gpa/di per

feet at applied pressure head for trenches with tiles as drainage material, and ranged from 0.0040 to 0.0044 for the two HIL without tiles. These values were determined for flow rates ranging from 0.15 to 0.25 gpm/ft. Use of these values for flows outside this range should be performed with caution until additional data is collected. The results reported here represent the conditions at one landfill. While landfill conditions vary from site to site, these results provide a starting point for the design of a HD-LHD.

CHAPTER 1
HORIZONTAL INJECTION LEACHATE RECYCLE:
CONSTRUCTION-OPERATION GUIDELINES AND DESIGN PROCEDURES

Introduction

This paper outlines guidelines for the design, construction, and operation of horizontal injection leachate recycle systems (HI-LRS) at solid waste landfills. Horizontal injection leachate recycle is a method to recirculate collected landfill leachate back through landfilled material using horizontal wells buried within the waste. Leachate recycle offers the potential to accelerate the decomposition of biodegradable components in the landfill, to provide some measure of leachate treatment, and to provide a method of leachate volume management.

The guidelines presented here are the result of three years of work in the construction and operation of a HI-LRS at a Florida landfill (Chapter 3) and an analysis of theoretical relationships of steady-state saturated flow from a horizontal injection well (Chapter 2). The design guidelines presented apply to the sizing and placement of the landfill's hydraulic structures, specifically the LRS pumping and conveyance system, the injection lines, and the leachate collection system. Other landfill aspects not considered here, but which play a potential role in the design of such a system include

the effect of leachate recycle application rates and frequency on the landfill decomposition process, the effect of the RI-LRI on the overall hydrologic balance at the site, and the collection of landfill gas from the injection lines. This chapter provides fundamental information for the implementation of RI-LRI at solid waste landfills.

Theory of Horizontal Injection

The principles behind the leachate recycle process have been well illustrated in pilot-scale work (Leslie et al. 1979, Bohland 1980) and have been applied at numerous full-scale sites in the United States and throughout the world (Petala and Andersen 1983, Berber and Morris 1984). A variety of leachate recycle methods have been employed, each with distinct advantages and disadvantages. One method which possesses a number of advantages over other leachate recycle methods is horizontal injection. This technology utilizes perforated pipe or drainage material placed in horizontal trenches buried within the landfill waste. The trenches are constructed during the operational years of the landfill while waste is deposited.

A RI-LRI permits the recirculation of leachate to the interior of the landfill without exposure to the atmosphere. Placement of injection wells at different locations and elevations allows control of the rate and frequency of leachate application to areas within the landfill. Horizontal

Injection wells present minimal interferences with landfill surface operations and vehicle traffic.

Construction and Operation Guidelines

The construction of a RI-LM requires advance planning and timing in regard to the deposition of solid waste. During the process of waste placement, a compacted layer of solid waste is often placed above the zone of an injection line immediately after it is constructed, therefore operations must be timed in coordination with the landfill filling sequence. Construction in the close proximity of landfill vehicular traffic may be necessary and proper safety provisions must be made. The length of time a trench is left open should be limited to reduce odor problems, as well as to comply with regulations regarding the daily application of cover material to solid waste.

The utility of an injection line and space as a hydraulic distribution structure should be considered during construction. In a landfill, layers of cover material are distributed throughout, and may act as preferential channels or hydraulic barriers, leading to uneven leachate distribution and surface shape. Large pressures often develop in the area surrounding a well, and if a layer of highly permeable material interrupts a zone of high pressure, a direct path to the surface, base, or bottom of the landfill may result. As waste is deposited around an injection well, direct contact with the well cover material should be minimized and RIL

should be surrounded by solid waste to the greatest extent possible.

Cover materials with hydraulic properties comparable to solid waste would provide the most uniform moisture distribution. The use of an alternative daily cover such as a temporary tarp or foam would create more uniform landfill conditions, and minimize potential seepage channels. Short circuiting may also result from proximity to a gas well. Use of vertical gas wells in areas of injection should be limited.

Differential settlement of landfilled material should be considered in selecting pipe materials. This is most important for the non-perforated pipe section connecting the EIL to the LEO pumping system. If a drainage material is provided in the trench, hydraulic flow would still be possible in the event of a pipe break. Provision must be made to connect the injection lines to a main distribution line along the boundary of the landfill. The ends of the injection branch should be located a sufficient distance from the landfill face to minimize potential seepage. A clay seal or flow collar should be placed between the perforated distribution and non-perforated connection pipe to prevent preferential flow along the non-perforated pipe. Flexible connections from the EIL to the main distribution line should provide allowance for settlement.

The type of landfill unit also impacts the construction and use of the RI-LSS. In-ground landfills, or landfills surrounded by large earthen berms, are less likely to pose a seepage problem than high-rise landfills. In-ground landfills could be operated at greater injection flow rates than high-rise landfills. As part of RI-LSS operation, the pressure and flow of the injection system should be monitored routinely during operation. Mechanisms should exist to adjust the rate of leachate flow to maintain performance standards within the design of the system. The landfill should be inspected routinely for seepage that might be the result of the injection system, so that adjustments to the RI-LSS operation can be made.

Design Elements

In the sizing and placement of a RI-LSS, features that must be considered include:

1. Maximizing the distribution of moisture throughout the landfill.
2. Minimizing the potential for leachate seepage in the surface or side slopes of the landfill.
3. Maintaining the design performance of the leachate collection system in regard to the leachate head above the liner.

To design a RI-LSS the pipes must be sized and spaced, the rate and frequency of flow application must be increased, and the leachate collection system must be designed to accommodate maximum design leachate recycle rates.

The volume and rate of leachate which will be recycled at a landfill is a function of the landfill's hydrology, topography, and configuration. The factors which play a large role in determining leachate recycle rates include:

1. Total volume of unrelated infiltrating into the landfill.
2. Volume of storm-water retained within the landfill with which must be managed as leachate.
3. Leachate treatment and disposal capacity within the landfill.
4. Leachate storage capacity at a landfill site.

The latter three factors are determined as part of the overall design process and should be evaluated in conjunction with the "SI-DBS" design. A hydrologic budget should be developed which includes the SI-DBS. Hydrologic models which incorporate rainfall, storm water runoff, and leachate recycle volumes (Schroeder et al. 1984, for example) should be employed to determine design leachate recycle rates.

Injection Line Spacing

The procedure outlined here for the sizing and spacing of the injection lines, utilizes equations which describe the saturated flow region surrounding a well pumped under pressure (Chapter 7). The equations were developed using a line source saturated flow solution and neglect the effect of capillary potential and dispersion. Under some operating conditions, a horizontal injection leachate recycle system at a landfill

would be operated on an intermittent basis and may not reach steady state conditions. In addition, capillary forces display a role in moisture transport as a landfill (Korfiatis et al., 1983). However, the method outlined here is not intended as a tool to model leachate production. The extended use of these equations is to establish threshold design conditions which allow the design of the RI-LRI to perform under conditions of maximum design flow.

A schematic of the saturated flow system developed around a RI-LRI is presented in Figure 4-1. The shape of the saturated zone may be described as

$$x = \frac{q}{2k_y} \tan^{-1} \left[\frac{z}{r} \sqrt{\frac{k_y}{k_x}} \right] \quad 4-1$$

where q is the linear flow rate, k_y is the vertical hydraulic conductivity, and k_x is the horizontal hydraulic conductivity. The maximum height above the RI may be described as

$$z_{\max} = \frac{q}{\alpha_p k_y k_x} \quad 4-2$$

The maximum spread of leachate may be described as

$$x_{\max} = \frac{q}{2k_y} \quad 4-3$$

The lateral distance of leachate spread at the well may be described as

$$x_{\max} = \frac{q}{2k_y} \quad 4-4$$

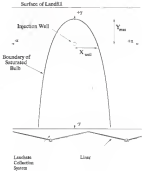


Figure 4-1. Injection Well Flow System

The above equations may be used to estimate the spacing of injection lines to minimize solute distribution under a given set of operating conditions. Spacing may be based upon selected zone coverage for average or peak flow conditions. The degree of overlap of zones may be selected by the design engineer based on the engineer's confidence in the design data and the flexibility of the pumping system. An additional distance of insoluble spread which accounts for the unactivated flow zone that would surround the activated zone could be incorporated. The equations can be used to determine locations and operating conditions to minimize the potential for seepage occurring.

Design of Pumping System

The pressure required to deliver a given flow rate in a specified injection system may be determined by use of empirical data for values of injection flow rate to applied pressure ratio, defined here as α , determined from other landfill sites. Typical values of α measured at the Alachua County Southwest Landfill (Chapter 3) ranged from 3.8 to 9,801 gpm/ft² per foot of applied pressure (water column). These measurements were made from tire-filled trenches with 3/4 inch PVC injection lines sparged at flow rates of 0.15 to 0.21 gpm/ft. For leaky lines without shrouded tires, α ranged from 0.21 to 2.46 gpm/ft per foot of applied pressure.

The pressure required to deliver a given flow rate in a specified injection system may also estimated from values for waste hydraulic conductivity using of the following equation:

$$P_h = \frac{Q}{2\pi b_p K_p} \ln \left(\frac{R}{r_w} \right) \quad 4-3$$

where P_h is the applied pressure head, r_w is the effective well radius, and R is the radius of influence. The radius of influence may be thought of as the distance beneath a horizontal line source at which the effect of line source becomes negligible.

Using these relationships, a system curve may be established from which a pump can be selected. The system curve for steady state injection into a well of length L , may be determined with the following equations:

$$P_h = L + \frac{V_p^2}{2g} + \frac{fL}{D} \frac{V_p^2}{2g} + (K F_p) \frac{V_p^2}{2g} + \frac{Q}{K} \quad 4-4$$

$$P_h = L + \frac{V_p^2}{2g} + \frac{fL}{D} \frac{V_p^2}{2g} + (K F_p) \frac{V_p^2}{2g} + \frac{Q}{2\pi b_p K_p} \ln \left(\frac{R}{r_w} \right) \quad 4-5$$

where V_p is the pipe velocity, Q is the injected flow rate, f is the friction factor, D is the internal pipe diameter, and $K F_p$ is the sum of the minor loss coefficients.

Solid Waste Properties

The hydraulic conductivity of the landfilled solid waste must be estimated to design the injection well system. The

hydraulic conductivity of MSW has been reported in a number of studies, and the values have been found to vary greatly. A summary of reported MSW hydraulic conductivities is presented in Table 4-1 and includes information regarding the vertical hydraulic conductivity determined at the MSWLF as a previous study (Fonseca et al. In press a). Also included in Table 4-1 are the estimated $(k_h k_v)^{1/2}$ values resulting from the horizontal injection experiment at MSWLF (Chapter 3).

Since hydraulic conductivity values are the greatest unknown in this process, the pumping system should be designed to allow flexible control over the injected pressures and flow rates over an expected range of conditions. As Lemarini is homogeneous and this uncertainty requires that some degree of operational control be available after the system is in place.

Leachate Collection System

Typical design of a leachate collection system (LCS) includes performing simulations with a hydrologic model to determine an infiltration rate for leachate into the leachate collection system drainage layer. From this, the head above the liner may be determined as a function of the characteristics of the LCS using a number of equations (Moore 1980, Mookerjee 1984). In the design of an EL-LCS, the equations developed for the slope of the saturated flow zone may be employed to estimate the infiltration of leachate into the LCS. This allows the effect of continuous injection at large pressures to be predicted.

Table 4-1. Solid Waste Hydraulic Conductivity

Source	Comment	Hydraulic Conductivity (cm/year)
Fungaroil and Geyssier (1979)	Laboratory Tests, Shredded Waste	$1 \times 10^{-5} - 2 \times 10^{-4}$
Garfield et al. (1984)	Laboratory Tests	$8 \times 10^{-5} - 2 \times 10^{-3}$
Gwin et al. (1985)	Laboratory Tests	1.5×10^{-4}
Gwin et al. (1988)	Landfill Pump Test	$1 \times 10^{-3} - 1.5 \times 10^{-3}$
Schroeder (1988)	HELE Model pore-size Values	2×10^{-4}
Tomson et al. (1990)	Landfill Tests, Infiltration from ACWL K_{eff}	$2 \times 10^{-4} - 4 \times 10^{-4}$
Shank (1993)	Landfill Pump Test ACWL K_{eff}	$2.7 \times 10^{-3} - 4 \times 10^{-3}$
This Study (CHAPTER 5)	Estimate from SI-LSS ACWL $(\frac{1}{2} K_{eff})^{1/2}$	$2.8 \times 10^{-4} - 2.8 \times 10^{-3}$

*Based on $K_s/K_{eff} = 100$

At an infinite distance beneath a horizontal injection well, the downward infiltration will equal the injected flow, or q_0/b_1 . The closer the LCR is located to the injection well, the greater the infiltration rate of leachate per unit area into the leachate collection system. This flow rate, or infiltration rate into the LCR, may be determined for the entire width of the secured zone using the following equation and Eqn. 4-3, to determine $K(y)$:

$$q(y) = \frac{2K_{\text{avg}}}{K(y)} \frac{q_0}{b_1} \quad 4-4$$

In the design of a LCR, simulations should be performed which predict leachate head above the liner resulting from site rainfall at times when the depth of water is small. The maximum conditions developed in the injection system should then be tested to determine if the limit design requirements are met.

Conclusions

A method was outlined to design a horizontal injection leachate recycle system. The design was based on the hydraulic sizing and spacing of the injection lines to maintain the distribution of leachate while minimizing potential for surface seepage. The procedure involved the use of steady-state saturated flow equations to approximate the spread of the secured zone surrounding an injection line.

The prediction of leachate generation rates and solute movement rates requires the use of more sophisticated numerical models, which account for the many complexities of a landfill.

Additional factors may affect the design and operation of a SI-LSI. The method outlined here should provide guidelines for SI-LSI design, and which should allow flexibility to examine various scenarios to meet specific goals in SI-LSI design. Additional work in the application of this technology to other sites will aid in the determination of the effects of the many additional factors that will ultimately require consideration.

CHAPTER 8 SUMMARY AND CONCLUSIONS

This dissertation reviewed the use of a technology known as horizontal injection leachate recycle. This technology utilizes horizontal wells constructed to the solid waste of a landfill to allow recirculation of leachate to the landfilled waste. The advantages of leachate recycle are many, and include the following.

1. Leachate recycle provides a means of hydrologic management at landfill sites.
2. The recirculation of leachate promotes the acceleration of biological stabilization of the solid waste.
3. Recycled leachate removes some treatment from the biologically active landfill.
4. A landfill which has been biologically stabilized presents a substantially reduced risk to the environment as a result of landfill operations.
5. Biological stabilization may obviate the need for landfill closure and may allow the incineration of stabilized material.

Horizontal injection is one of several methods of leachate recycle, and the advantages of this method include

1. Leachate may be recycled to the landfill without direct exposure to the atmosphere.
2. The point of leachate application may be controlled, allowing the fluidity to target specific areas and

levels within the landfill, as well as the rate and frequency of application.

1. Horizontal injection wells have little impact on daily landfill operations in regard to surface area requirements and the movement of landfill vehicles and equipment.

2. Horizontal injection wells allow the addition of various chemical or biological additives to the landfill.

The major limitation for the application of horizontal injection at full-scale landfills is the lack of full-scale operational experience to provide the design, construction, and operation guidelines. This study was designed to provide such information.

A method of analyzing the predicted zones of saturated flow resulting from the injection of leachate into a porous media was analyzed by treating the injection line as a line source. To facilitate the solution in the form of simple analytical equations, capillary forces and dispersion were neglected. The resulting solution was useful for the predicting flow boundaries for various flow conditions which could be used in design, but was not intended to model leachate solute transport in complex systems. Equations were developed for isotropic and anisotropic media to determine the steady state zones of saturation surrounding an injection well resulting from injections at a given flow rate.

This study reviewed the construction and operation of a HI-LRS at a composite lined, operating landfill in Florida. Construction of Injection Lines at the Florida Landfill began

in 1970, and the first injection began in 1973. Through August 1974, a total of 7,000,000 gallons of leachate were recirculated into 11 injection lines. The system was successful in providing a means of recycling large volumes of leachate to the landfill. The most dramatic problem associated with the operation of the RI-LRS was the short-circuiting of leachate to the surface through permeable zones in the landfill. Typical leachate flow rates ranged from 2,044 to 2,857 gpm/ft. per foot of applied head at a flow range of 1.5 to 2.2 gpm/ft. Injection flow did decrease over time to a limited extent.

This document outlines guidelines for the design, construction, and operation of RI-LRS at full-scale landfills. The information gathered in the field research project and the theoretical flow equations were used to outline a preliminary design technique and procedure. The technique was developed to allow estimates for the spacing of injection lines, placement and operation of lines to preclude the escape of leachate to the surface, and the placement and operation of lines for the design of leachate collection systems. The procedures outlined are proposed as a method for the design and operation of a RI LRS which in turn will allow additional field data to be gathered for the refinement of the design of such systems.

The operation and design of horizontal injertive leachate recycle systems were reviewed in regard to the hydraulic

design of the landfill components (LFG piping and pump sizing, LFG design). Additional areas that influence the design of an injection system may include:

1. The effect of leachate application on biological activity of the landfill microorganisms.
2. The hydrology of the entire landfill site.
3. Use of the Sile to collect landfill gas.
4. The effect of leachate recycle on leachate quality.

Additional field studies at the site discussed here and other sites is necessary for further development of this system.

APPENDIX B
LEAKAGE INJECTION GRAPHS

This appendix presents graphs of the data collected for the horizontal injection line tests at LOML. The results cover injection lines I through 11, and cover a period from February 1983 through August 1984. The following graphs are presented for each injection line:

I. Injection flow and pressure versus calendar time (day 1 represents January 1, 1983).

II. Injection flow and pressure versus pumping time.

III. Linear flow rate and total energy head at entrance of injection line versus pumping time.

IV. Linear flow rate/total energy head ratio (α) versus pumping time.

Note: In the data in graphs of pumping time and cumulative pumped volume represents injection tests where data logger files were lost. Total volume and pumping time were accounted for in these cases.

Legend: The following graphs were formatted so flow is reported by the thicker solid line and pressure by the thinner dotted line. The flow is generally scaled higher than the pressure.

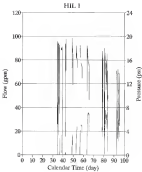


Figure A-1. Injection Flow Rate and Pressure versus Calendar Time: HIL 1 (February 1992 - April 1993)

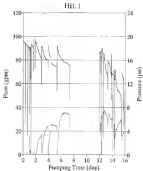


Figure A-2. Injection Flow Rate and Pressure versus Injection Time HIL 1 (February 1982 - April 1985)

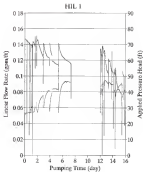


Figure 4-3. Linear Flow Rate and Total Applied Head versus Injection Time WEL 1 (February 1993 - April 1993)

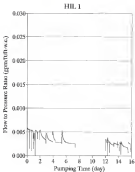


Figure A-4. a versus Injection Time HIL 1
(February 1993 - April 1993)

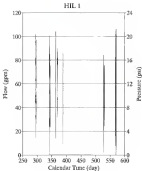


Figure A-5. Injection Flow Rate and Pressure versus Calendar Time HIL 1 (September 1953 - August 1964)

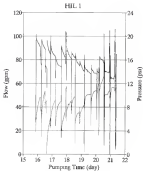
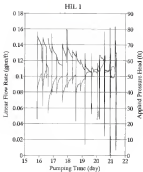


Figure A-4 Injection Flow Rate and Pressure versus Injection Time HIL 1 (September 1993 - August 1994)



Figures 8-9. Linear Flow Rate and Total Applied Head versus Sejection Time: HIL 1 (September 1982 - August 1984)

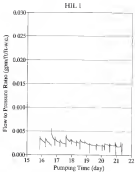


Figure A-6. α versus Injection time; HIL 1
(September 1993 - August 1994)

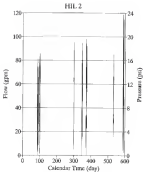


Figure A-8. Injection Flow Rate and Pressure versus Calendar Time: HIL 2

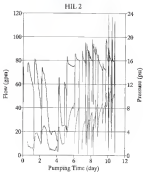


Figure A-13. Injection Flow Rate and Pressure versus Injection Time HIL 2

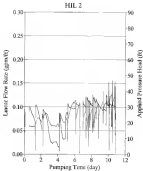


Figure 8-15. Linear Flow Rate and Total Applied Head versus Deposition Time: HIL 2

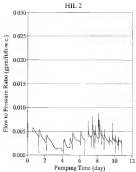


Figure A-11. α versus injection time: HIL 2

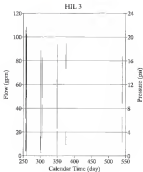


Figure B-13. Injection Flow Rate and Pressure versus Calendar Time--HIL 3

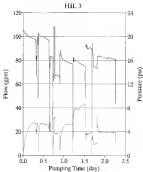


Figure A-14. Injection Flow Rate and Pressure versus Injection Time, HIL 3

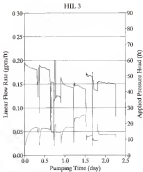


Figure A-18. Linear Flow Rate and Total Applied Head versus Injection Time HIL 3

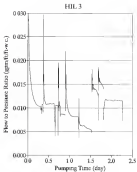


Figure A-14. a various Injection Times HIL 3

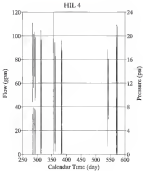


Figure B-17: Injection Flow Rate and Pressure versus Calendar Time, HIL 4

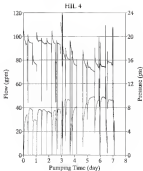


Figure A-18. Injection Flow Rate and Pressure versus Injection Time: HIL 4

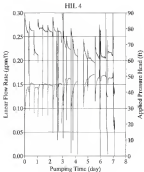


Figure A-14. Linear Flow Rate and Total Applied head versus injection time; HIL 4

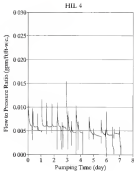


Figure A-20. q versus injection time: HIL 4

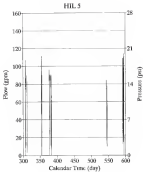
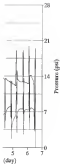


Figure 2-21. Injection Flow Rate and Pressure versus Calendar Time HIL 5



Pressure versus
Time

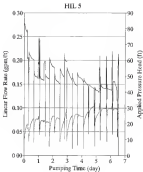


Figure A-23. Linear Flow Rate and Total Applied Head versus Injection Time: HIL 5

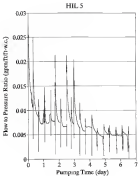


Figure A-28. α versus Injection Time HIL 5

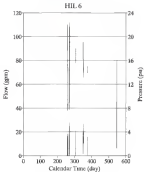


Figure A-25 Injection Flow Rate and Pressure versus Calendar Time, HIL 6

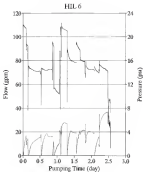


Figure A-18 Injection Flow Rate and Pressure versus
Injection Time: HIL 6

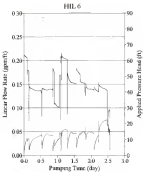


Figure A-29. Linear Flow Rate and Total Applied Head versus Injection Time; HIL-6

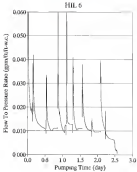


Figure A-24. q versus Injection times and q

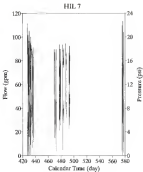


Figure A-29. Injection Flow Rate and Pressure versus Calendar Time: HIL 7

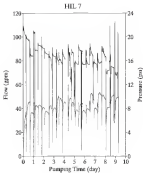


Figure A-39. Injection Flow Rate and Pressure versus Injection Time: HIL 7

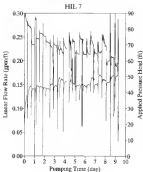


Figure A-33. Linear Flow Rate and Total Applied Head versus Injection Time: HIL 7

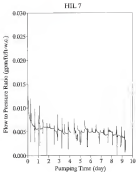


Figure 3-12. "a versus Injection Time" HIL 7

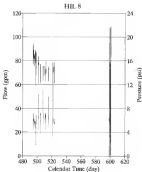


Figure A-23: Collection Flow Rate and Pressure versus Calendar Time: HIL 8

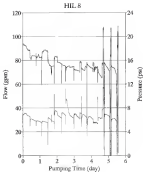


Figure 4-14: Injecting Flow Rate and Pressure versus Injection Time, HIL 8

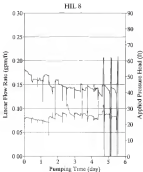


Figure A-25. Linear Flow Rate and Total Applied Head versus Injection Time HIL 8

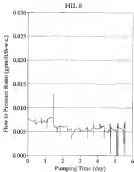


Figure A-38. A versus Injection Time HIL 8

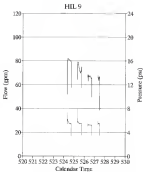


Figure A-9F. Injection Flow Rate and Pressure versus Calendar Time 5/5, 9

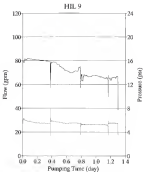


Figure A-38. Injection Flow Rate and Pressure versus Injection Time; well 9

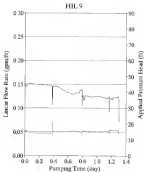


Figure A-99 Linear Flow Rate and Total Applied Head versus Injection Time HIL 9

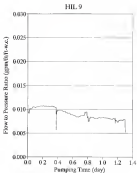


Figure A-10. A minimum Injection Time: HIL 9

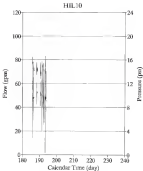


Figure 4-41. Injection Flow Rate and Pressure versus Calendar Time: HIL 10

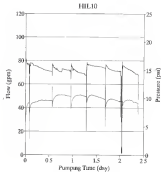


Figure A-41. Injection Flow Rate and Pressure versus Injection Time HIL 10

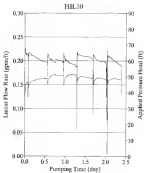


Figure B-13 Linear Flow Rate and Total Applied Head versus Injection Time, HIL 10

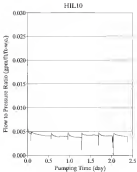


Figure A-44. α versus Injection Time: HIL 10

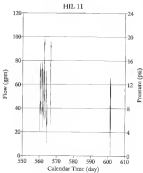


Figure 2-18. Injection Flow Rate and Pressure versus Calendar Time; HIL-11

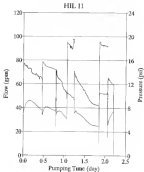


Figure A-46. Injection Flow Rate and Pressure versus Injection Time—HIL 11

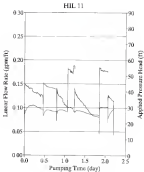


Figure B-67: Linear Flow Rate and Total Applied Head versus Injection Time HIL 11

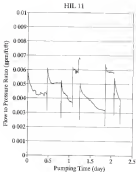


Figure A-48. Q versus Injection Time: HIL 11

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BIOGRAPHICAL SKETCH

Timothy W. Townsend was born in San Diego, California, on October 17, 1947, to Timothy A. and Cynthia Townsend. He graduated from Orange Park High School in Orange Park, Florida, in 1966.

He enrolled in the University of Florida in September, 1966, and graduated with high honors with a Bachelor of Science degree from the Department of Environmental Engineering Science in December 1969. He enrolled in graduate school in the same department in January, 1970, to study solid and hazardous waste. He graduated with his Master of Engineering degree in December, 1972.

Tim was married to Marlene Elizabeth on March 14, 1970.

I certify that I have read this study and that in my opinion it conforms to acceptable standards of scholarly presentation and is fully adequate, in scope and quality, as a dissertation for the degree of Doctor of Philosophy.


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This dissertation was submitted to the Graduate Faculty of the College of Engineering and to the Graduate School and was accepted as partial fulfillment of the requirements for the degree of Doctor of Philosophy.

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